



# The Utility of Biological Monitoring for Manganese in Ferroalloy Smelter Workers in South Africa

Jonathan E. Myers<sup>1,\*</sup>, Mary Lou Thompson<sup>1,3</sup>, Inakshi Naik<sup>2</sup>, Penny Theodorou<sup>2</sup>,  
Eric Esswein<sup>2,4</sup>, Halina Tassell<sup>2</sup>, Aarti Daya<sup>2</sup>, Kevin Renton<sup>2</sup>, Adri Spies<sup>2</sup>,  
Janice Paicker<sup>5</sup>, Taryn Young<sup>1</sup>, Mohamed Jeebhay<sup>1</sup>, Suzan Ramushu<sup>1</sup>,  
Leslie London<sup>1</sup>, David J. Rees<sup>2</sup>

<sup>1</sup>Occupational and Environmental Health Research Unit, School of Public Health and Primary Health Care,  
Faculty of Health Sciences, University of Cape Town, Anzio Road Observatory, 7925 Cape Town, South Africa

<sup>2</sup>National Centre for Occupational Health, South Africa

<sup>3</sup>Department of Biostatistics, University of Washington, Seattle, WA, USA

<sup>4</sup>National Institute for Occupational Safety and Health, USA

<sup>5</sup>South African Institute for Medical Research, South Africa

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## Abstract

Five hundred and nine workers at a manganese (Mn) smelting works comprising eight production facilities and 67 external controls were studied cross-sectionally. Exposure measures from personal sampling included inhalable dust, cumulative exposure indices (CEI) and average intensity (INT = CEI/years exposed) calculated for the current job at the smelter and also across all jobs held by subjects. Biological exposure was measured by Mn in the blood (MnB) and urine (MnU) and biological effect was measured by serum prolactin. Average lifetime exposure intensity across all jobs ranged from near 0 (0.06  $\mu\text{g}/\text{m}^3$ ) for unexposed external referents to 5  $\text{mg}/\text{m}^3$ . Atmospheric exposures and MnB and MnU distributions were consistent with published data for both unexposed and smelter workers. Associations between biological exposures and groups defined by atmospheric exposures in the current job were substantial for MnB, less so for MnU and absent for serum prolactin. Random sampling of MnB measurements representative of a group of workers with more than 1–2 years of service in the same job and notionally homogenous exposure conditions could serve as a cross-sectional predictor of atmospheric Mn exposure in the current job, as well as for surveillance of Mn exposure trends over time. Correlations at the individual level were only modest for MnB (33% of the variance in log atmospheric Mn intensity in the current job was explained by log MnB), much worse for MnU (only 7%). However, a receiver operating characteristic (ROC) analysis was performed which showed that it is possible to use a MnB cut-off of 10  $\mu\text{g}/\text{l}$  (the 95th percentile in the unexposed) to good effect as a screening tool to discriminate between individual exposures exceeding and falling below a relatively strict atmospheric Mn exposure threshold at the ACGIH threshold limit value (TLV) of 0.2  $\text{mg}/\text{m}^3$ . MnU has no utility as a measure of biological exposure nor does serum prolactin as a measure of biological effect.

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## INTRODUCTION

This study of atmospheric and biological exposures in relation to neurobehavioural outcomes among large numbers of ferroalloy workers aimed to shed further

\* Corresponding author. Tel.: +27-21-4066306;  
fax: +27-21-4066163.  
E-mail address: myers@cormack.uct.ac.za (J.E. Myers).

light on associations with and between biological exposures (manganese concentrations in blood (MnB) and urine (MnU)) and biological effect (serum prolactin) measures and exposure intensity and cumulative exposure for manganese based on either the current job or across all exposed jobs. The potential of these biological measures for group surveillance, and for individual screening in relation to the exceedances of commonly used exposure limits such as the ACGIH threshold limit value (TLV), was also explored.

There are some uncertainties about the value of biological exposure or biological effect monitoring in the estimation of workplace exposure (which can be difficult, expensive and impractical to determine) and exposure-related effects (Smargiassi and Mutti, 1999). Blood and urine Mn have been found to be inconsistently related to atmospheric Mn (MnA) exposure (whether measured as exposure intensity or as cumulative exposure either in the current job or across all exposed jobs), to each other; and also to biological effect measures and other (e.g. neurobehavioural) outcomes.

This applies both to group comparisons as well as to individually based correlations. Roels et al. (1987) found that in chemical plant workers MnB was not associated with MnA on a group basis, and that there was neither correlation between MnB and MnU, nor between either of these and atmospheric Mn exposure at the individual level. In a 1992 study of battery workers exposed to MnO<sub>2</sub>, Roels et al. found that only MnU, but neither MnB nor serum prolactin, was significantly associated with MnA on a group basis. They found no relationship between MnB or MnU and atmospheric exposures measured as duration, current intensity or cumulative exposure at the individual level. Only cumulative exposure to atmospheric Mn was associated with neurobehavioural abnormalities.

On the other hand, Lucchini et al. (1995) found significant associations between MnB and MnU, and between both of these and the cumulative exposure index (CEI), at the individual level in ferroalloy workers who had been temporarily laid off. Furthermore, MnB and MnU were both associated with neurobehavioural abnormalities. As these relationships, especially for MnB, were strengthened after cessation of exposure, MnB was thought to reflect the Mn body burden more accurately without confounding by acute exposure. By contrast, for currently employed workers, Lucchini et al. (1999) found no relationship between biological exposure measures and CEI on the one hand, nor with neurobehavioural test results on the other. There was, however, an association between MnB and exposure intensity.

More recently, Apostoli et al. (2000) investigated the suitability of MnB and MnU for exposure assessment, and concluded that while MnB and MnU can discriminate between exposed and unexposed groups of workers, and while there was a linear relationship between MnB and atmospheric Mn intensity in the current job, variability was too high (with only 13% of the variance explained) to be of use in individual biological monitoring.

Roels et al. (1992) found that serum prolactin as a presumptive measure of effect was not related to atmospheric Mn exposure, while Mutti et al. (1996) and Smargiassi and Mutti (1999) found raised levels in ferroalloy workers. Mutti et al. (1996) also found that MnB and MnU were associated with serum prolactin levels.

Occupational health practitioners responsible for Mn exposed workers require surveillance tools to help reduce exposure and prevent adverse health effects. The utility of biological exposure and effect measures has become increasingly important to establish as occupational and environmental exposure limits for manganese have been set increasingly lower in recent years in order to prevent early nervous system effects. Exposure–response relationships with neurobehavioural outcomes are dealt with in a companion paper (Myers et al., in press).

## MATERIALS AND METHODS

A cross-sectional study was conducted on 509 manganese exposed subjects drawn from one of eight production environments in a Mn smelting works in South Africa. Different plants or activities at the works were divided into notionally high (three ferro- and silico-Mn smelters), medium (one ferro-silicon smelter, raw and finished materials handling plants) and low (quality control laboratories, administration and security workers, and a chemical plant making no use of Mn). The low exposure group served as an internal control with no direct Mn exposure. The smelter has been producing Mn ferroalloys for about 50 years and is one of the largest global producers of Mn. Ore comes in by rail to a raw materials handling yard where it is crushed and sorted, and conveyed to the smelter plant furnaces where it is smelted using the Soderberg process. Molten ore is then tapped from the bottom of the furnaces and poured into ladles from where it is separated from slag and transported to casting bays. When the casts have cooled they are transported by front end loader to the final products yard where they

are crushed, sorted and dispatched. The two non-manganese producing plants at the works are a dense ferro-silicon smelter and a chemical factory producing phosphates. Apart from water misters in the materials handling yards there were no engineering controls limiting exposure to workers in place, particularly not at the furnaces. Where personal protective equipment was in use this involved disposable filtering face piece respirators. Few changes affecting occupational hygiene had been made over the years. Additionally, 67 external unexposed reference workers from an electrical fittings assembly plant were included in the study.

There were 1380 production workers at the works of which 200 subjects were randomly selected in each of three subjectively determined exposure groups. Maintenance workers with highly variable exposures were excluded. High exposure was considered to be above 2 mg/m<sup>3</sup>, medium between 2 and 0.1 mg/m<sup>3</sup>, and low exposure below 0.1 mg/m<sup>3</sup>.

Sampling for atmospheric Mn measurements followed a NIOSH method (NIOSH, 1977) to capture at least one individual from the highest 10% of an exposure group with 90% confidence. Subjects were drawn from all homogeneous exposure locations and/or jobs in smelter and control plants.

Full-shift personal breathing zone inhalable dust and fume samples were collected. Employees wore Gilian<sup>®</sup> Gilair constant-flow personal air sampling pumps using Institute of Medicine inhalable dust sampling heads with 25 mm × 1.2 µm pore size, mixed cellulose-ester membrane filters, connected to pumps by Tyvek<sup>®</sup> tubing. Pumps operated at 2 l/min. Blood and urine specimens were collected for the determination of MnB, MnU and serum prolactin as described elsewhere (Myers et al., in press).

All air samples were analysed by using a modified NIOSH method 7300, which was designed to optimise for the presence of Mn including insoluble SiMn (NIOSH, 1999). The IOM cassettes were first rinsed with deionised water to remove any dust, which might have remained in the cowl of the IOM filter cassette holder. The rinsate was added to the filter samples. Filters were digested using a CEM MARSX microwave digester. A combination of hydrochloric, nitric and hydrofluoric acid were used together to digest the Mn and any silica compounds, which might have bound to Mn compounds. A Varian<sup>®</sup> Vista simultaneous inductively coupled plasma optical emission spectrometer (ICP-OES) was used for all the analyses.

Three levels of in-house quality control were prepared by spiking blank filters with stock manganese solution. The filters were digested and analysed in the same

manner as the samples with each batch of analyses. The mean recovery was 102%. The CV ranged between 3.41 and 3.48%. Three levels of Certified Reference Material (CRM) from United States National Institute of Standards and Technology no. 2676d were digested and analysed in the same manner. Mean recovery was 101.6% (standard acceptable range 95–105%).

Heparinised whole blood samples were diluted 10 times using Triton X100 and manganese content on the diluent was measured by a Varian<sup>®</sup> graphite furnace atomic absorption spectrophotometer (Varian<sup>®</sup> Spectra AA 30) coupled with a Zeeman effect background correction system (Varian<sup>®</sup> Zeeman graphite tube atomiser). Standards and CRM (Seronorm<sup>™</sup> trace elements, whole blood, level II) were run at the beginning, during and at the end of each batch of samples that were analysed. The mean and the CV of all the results obtained from the CRM was 11.0 (10.3–11.7) µg/l (recommended value 13.0 µg/l) and 6.45%, respectively.

Urine samples were diluted five times with 0.1N nitric acid and analysed by a Varian<sup>®</sup> AA 975 atomic absorption spectrophotometer equipped with GTA-95 furnace and autosampler. Standards and CRM (Seronorm<sup>™</sup> trace elements, urine and BIO RAD Lyphocheck<sup>®</sup> level I) was run at the beginning, during and at the end of each batch of samples that were analysed. The mean and the CV of all the results obtained from the Seronorm R was 10.51 (8.23–11.8) µg/l (recommended value 13.0 µg/l) and 12.2%, respectively. The mean and the CV for BIO RAD Lyphocheck<sup>®</sup> level I was 6.3 (5.4–7.2) µg/l (recommended 5.9 (4.4–7.0) µg/l) and 14%, respectively.

Two different types of exposure intensity and cumulative exposure measures were calculated, reflecting lifetime and current job exposure. For all jobs in a working lifetime, a cumulative exposure index was calculated for each subject by multiplying the mean Mn inhalable dust concentration characteristic of each job or activity by the number of years worked in that job, and summing these products over all jobs worked by each subject at the works. The CEI was divided by total years of service at the works (LOS) to yield a measure of mean exposure intensity over all jobs at the works (INT). The same was done for exposures in the current job.

Uni- and bivariate data exploration were used to describe the atmospheric and biological exposure variables and explore their inter-relationships. The non-parametric distributions of MnB, MnU and serum prolactin were examined within categories of different atmospheric exposure variables including exposure

250 intensity and cumulative exposure in the current job  
251 and across all jobs held, as well as with total years of  
252 service at the smelter. Conclusions about the nature of  
253 associations between biological and atmospheric expo-  
254 sures at the grouped level are based on examination of  
255 coefficients in categorical exposure modelling using  
256 multiple regression. Individual correlations of log-  
257 transformed data were examined using Pearson's *r*  
258 correlation coefficient.

259 To address the question of whether MnB or MnU  
260 might be used as a surrogate for job history based  
261 exposure measures to flag individuals exceeding cer-  
262 tain average exposure intensity levels, a receiver oper-  
263 ating characteristic (ROC) analysis was carried out.  
264 Stata 6.0 software was used (STATA, 1999). Associa-  
265 tions of biological exposure and effect measures with  
266 neurobehavioural outcomes are examined elsewhere  
267 (Myers et al., in press).

268 The study was approved by the ethics and research  
269 committee of the Health Sciences Faculty of the Uni-  
270 versity of Cape Town. Informed consent was signed by  
271 all participants. A Research Reference Panel was set up  
272 with representatives of workers and their trade unions,  
273 management and researchers to oversee all aspects of  
274 the study and to assist the research team. The reference  
275 panel served as a conduit for stakeholder input to the  
276 research process. Although the study was funded by the  
277 company, independence of the researchers in the  
278 design, conduct, analysis and interpretation and report-  
279 ing of the results was ensured in the research contract  
280 which included the right to presentation and publica-  
281 tion of findings in the scientific media.

## RESULTS

283 Some 442 personal inhalable dust samples were  
284 measured in various homogeneous exposure zones

Table 1  
Categorisation of exposure variables

Category	Exposure range (mg/m <sup>3</sup> )	Significance	<i>n</i>
Average exposure intensity across all jobs (mg/m <sup>3</sup> )			
0	0	Unexposed controls	67
1	0 < <i>x</i> ≤ 0.1	LOAEL <sup>a</sup> in literature	105
2	0.1 < <i>x</i> ≤ 0.2	ACGIH TLV (1996)	50
3	0.2 < <i>x</i> ≤ 1	SA OEL <sup>b</sup> for fumes	235
4	1 < <i>x</i> ≤ 2	Company advisors' safe level	59
5	>2		59
Cumulative exposure index (mg-years/m <sup>3</sup> )			
0	0		67
1	0 < <i>x</i> ≤ 1.3		104
2	1.3 < <i>x</i> ≤ 5.4		98
3	5.4 < <i>x</i> ≤ 10.6		103
4	10.6 < <i>x</i> ≤ 22.4		101
5	>22.4		102

<sup>a</sup> Lowest observable adverse effect level.

<sup>b</sup> South Africa occupational exposure limit.

for jobs in different production locations enabling  
the construction of exposure indices shown in Table 1.

Arithmetic means, geometric means and maxima  
for intensity of exposure in the current job were  
0.006, 0.005 and 0.01 within the chemical plant;  
0.04, 0.03 and 0.11 for security; 0.04, 0.03 and  
0.04 for administrative personnel; 0.21, 0.16 and  
0.52 for the ferri-silicon smelter; and 0.36, 0.26  
and 0.56 for materials handling. The three Mn smelter  
plants had exposures ranging from a low of 0.80, 0.69  
and 1.29 through 1.40, 1.16 and 2.82 to a high of 2.70,  
1.97 and 5.08.

No association was found with serum prolactin and  
any measure of atmospheric or biological exposure  
(Table 2). With one exception, 27 µg/l, all values fell  
within the normal laboratory range of 2.1–17.7 µg/l for  
males. Fig. 1 typifies the absence of association in the

Table 2  
Atmospheric and biological exposures

	<i>n</i> <sup>a</sup>	Mean (S.D.)	Geometric mean (G.S.D.)	Minimum	Maximum
Cumulative exposure index across all jobs (mg-years/m <sup>3</sup> )	508	16.0 (22.4)	5.1 (6.7)	0	137.6
CEI in current job only	511	12.7 (21.3)	3.3 (7.9)	0	137.6
Average intensity across all jobs (mg/m <sup>3</sup> )		0.8 (1.1)	0.3 (5.5)	0	5.1
INT for the current job only	507	0.9 (1.2)	0.3 (6.3)	0.003	5.1
Total length of service	576	17.2 (8.1)	14.7 (2.0)	0.4	42
MnB (µg/l)	482	11.7 (5.6)	10.6 (1.6)	3.3	44
MnU (µg/l)	481	9.2 (19.1)	3.3 (3.9)	0.5	170
Serum prolactin (µg/l)	456	6.1 (2.9)	5.5 (1.5)	1.7	27

<sup>a</sup> For workers in the smelter works only.

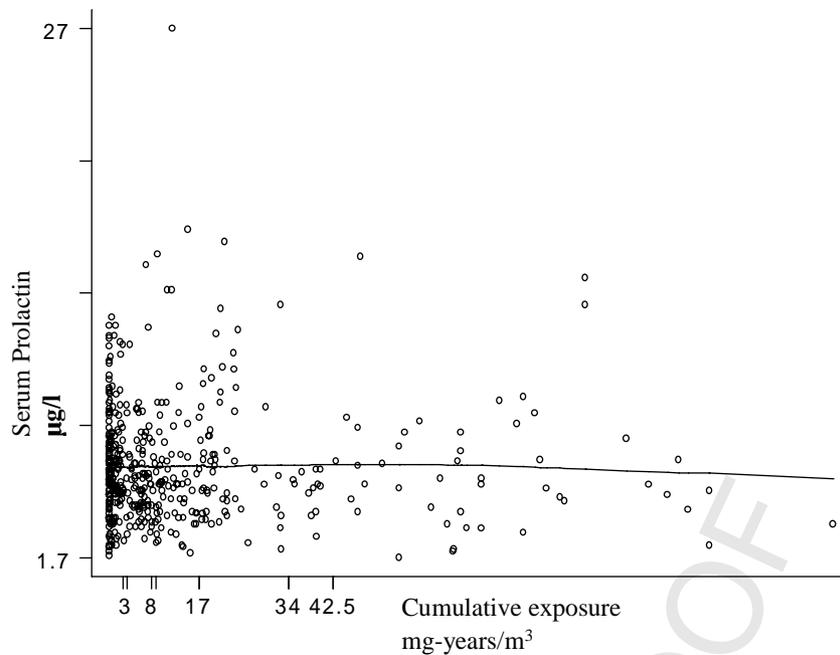


Fig. 1. Smoothed bivariate plot of serum prolactin vs. cumulative exposure across all jobs.

302 case of cumulative exposure across all jobs. The results  
303 for exposure in the current job are similar.

304 Table 3 shows non-parametric distributions for MnB  
305 by different exposure categories. The median values  
306 are very close to the geometric means 6.2, 7.5, 10, 11.1,  
307 14.2 and 15.2, respectively. Multiple regression ana-  
308 lysis shows that mean log MnB in every exposed  
309 category is significantly different from the unexposed,  
310 and also increases significantly for each higher cate-  
311 gory of current job exposure intensity amongst the  
312 exposed. The same is not true for average intensity  
313 across all exposed jobs in that the mean does not  
314 always increase significantly for each consecutive  
315 higher exposure category.

316 Fig. 2 shows the distribution of blood Mn by differ-  
317 ent exposure categories. Multiple regression analysis  
318 shows that mean log MnU for any category of atmo-

spheric Mn exposure is significantly higher than in the 319  
unexposed, but that there is no significant increase in 320  
the geometric mean between any two exposure cate- 321  
gories (Table 4). 322

323 Table 5 shows that individual correlations with 323  
logged exposure intensity in the current job are rela- 324  
tively good for log MnB. Fig. 3 shows some saturation 325  
at 2 mg/m<sup>3</sup>. Individual correlations with log exposure 326  
intensity in the current job are relatively poor with log 327  
MnU. The log MnB and log MnU are significantly 328  
correlated ( $r = 0.43$ ,  $P < 0.0001$ ). 329

330 Fig. 4 shows the estimated ROC curve for MnB used 330  
as a screen to identify individuals whose exposure 331  
intensity in the current job exceeds the ACGIH TLV 332  
level of 0.2 mg/m<sup>3</sup>. for exposure intensity in the current 333  
job. The discrimination here is quite good. For 334  
instance, if a threshold for MnB of 10 µg/l is used 335

Table 3  
MnB vs. current job exposure intensity

Intensity in current job (mg/m <sup>3</sup> )	MnB (µg/l)					
	<i>n</i>	Minimum	Maximum	Median	10th percentile	90th percentile
Unexposed	63	3.3	10.9	6.2	4.7	9.1
Total exposed	419	3.3	44	11.4	7	19.4
Intensity range						
0 < <i>x</i> ≤ 0.1	64	3.3	19.3	7.6	5.1	10.7
0.1 < <i>x</i> ≤ 0.2	67	4.6	41.2	9.4	6.4	14.7
0.2 < <i>x</i> ≤ 1.0	127	5.1	44	11.7	7.4	16.3
1.0 < <i>x</i> ≤ 2.0	67	8.2	43.3	13.1	9.5	21.0
>2.0	92	5.5	38.7	14.8	10.4	24.3

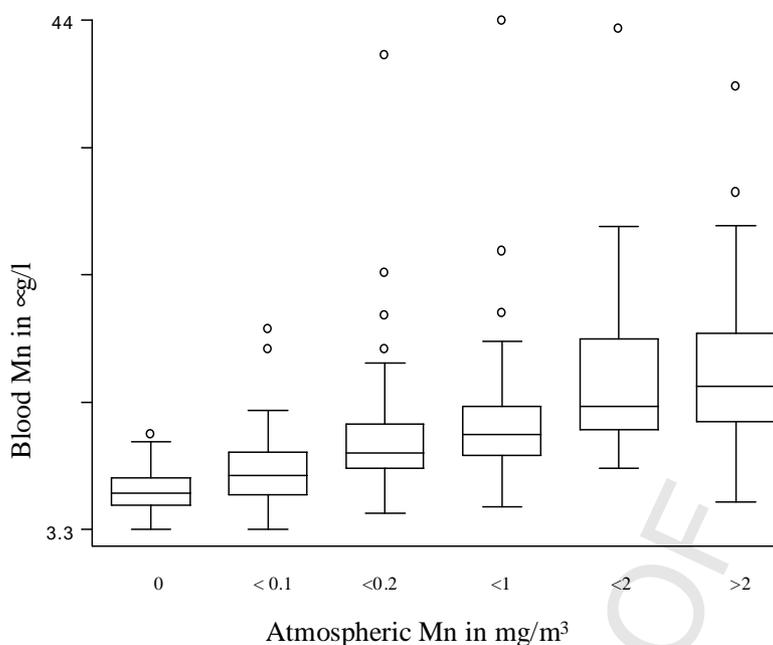


Fig. 2. MnB vs. category of exposure intensity in the current job.

Table 4  
MnU vs. current job exposure intensity

Intensity in current job (mg/m <sup>3</sup> )	MnU (µg/l)					
	<i>n</i>	Minimum	Maximum	Median	10th percentile	90th percentile
Unexposed	66	0.5	35	0.7	0.5	1.7
Total exposed	419	0.5	170	4	0.8	23.4
Intensity range						
0 < <i>x</i> ≤ 0.1	64	0.5	51	1.4	0.6	16
0.1 < <i>x</i> ≤ 0.2	64	0.5	98.7	3	0.6	17.9
0.2 < <i>x</i> ≤ 1.0	130	0.5	124	4.1	0.9	17.4
1.0 < <i>x</i> ≤ 2.0	64	0.6	48.4	6.2	1.3	26.4
>2.0	90	0.7	170	5.7	1.4	55.4

336 (based on the 95th percentile in the unexposed referent  
337 group) to discriminate, the estimated sensitivity (per-  
338 cent of those truly in exceedance of the ACGIH thresh-  
339 old who are correctly identified) is 80% (95%  
340 confidence interval: (75, 85%)); and estimated speci-  
341 ficity (percent of those truly below the ACGIH thresh-

old who are correctly identified) is 81% (95%  
confidence interval: (76, 86%)). The estimated positive  
predictive value (percent of those flagged by blood Mn  
who are truly in exceedance of the ACGIH threshold) is  
87% (95% confidence interval: (82, 92%)). The ROC  
curve for average exposure intensity across all jobs is  
similar but the performance is somewhat poorer.

Table 6 shows the results of a sensitivity analysis  
with estimated sensitivity, specificity, positive and  
negative predictive values and overall percent correctly  
identified for a blood Mn threshold of 10 µg/l used to  
identify individuals with exposure intensity levels in  
the current job and also average exposure intensity  
across all jobs in exceedance of 0.2, 0.5, 1 and 2 mg/  
m<sup>3</sup>, respectively. Performance for MnU was much  
poorer, and similar analyses with selected neurobehav-  
ioural outcomes performed even worse.

Table 5  
Individual correlations (Pearson's *r*) between logged values for  
MnB and MnU and logged environmental Mn

Atmospheric Mn	MnB	MnU
Mean in current job	0.57	0.26
CEI in current job	0.53	0.25
Mean across all jobs	0.44	0.23
CEI across all job	0.44	0.20
Length of service	0.27	0.16

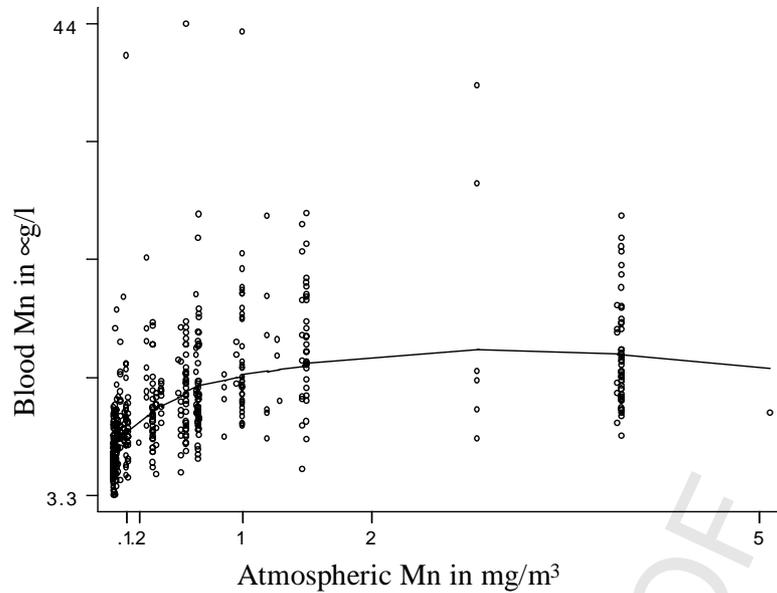
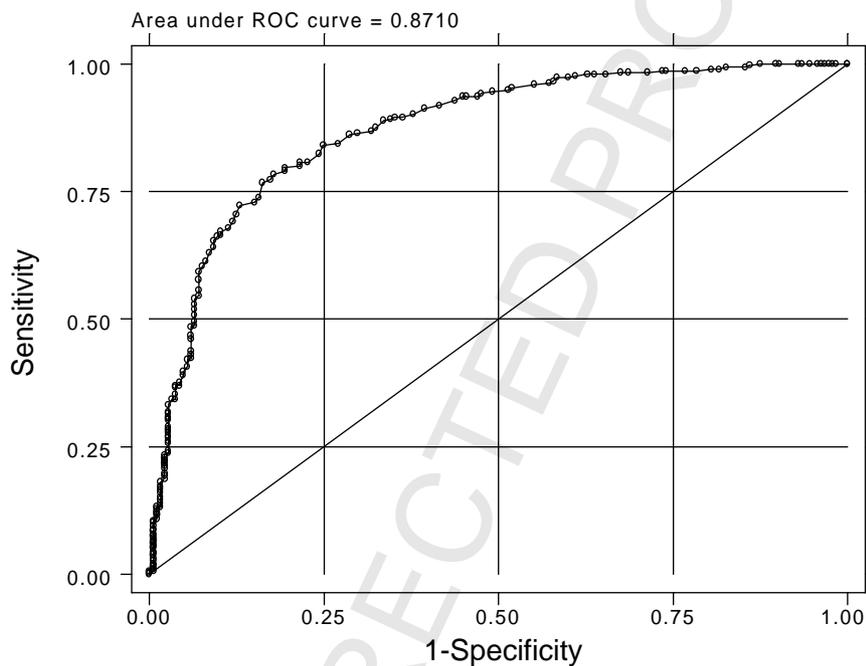


Fig. 3. MnB vs. current job exposure intensity.

Fig. 4. ROC analysis for MnB screening for exposure intensity in the current job exceeding the ACGIH TLV ( $0.2 \text{ mg/m}^3$ ).

## DISCUSSION AND CONCLUSIONS

360 These findings contribute to the inconsistent picture in  
 361 the literature. Overall, the distributions of our biomar-  
 362 kers were remarkably similar to those reported by  
 363 Apostoli et al. (2000) whose ferroalloy workers had  
 364 similar exposures. Serum prolactin was not associated  
 365 with any measure of atmospheric or biological exposure,  
 366 nor with neurobehavioural test results (Myers et al., in  
 367 press). This agrees with Roels et al. (1992) but not Mutti

et al. (1996) who found values of 9.77 (S.D. = 1.69) in 368  
 exposed ferroalloy workers and 4.65 (S.D. = 1.78) in 369  
 controls. The latter also found associations with MnB 370  
 and MnU. Smargiassi and Mutti (1999) later reported a 371  
 positive association between serum prolactin and atmo- 372  
 spheric Mn at low exposures. Serum prolactin was in our 373  
 study, however, negatively associated with smoking 374  
 status, as has been previously reported. 375

Significant associations at the group level were 376  
 found between both log MnB and log MnU and any 377

Table 6  
Performance of MnB in screening for atmospheric exposure intensity at different thresholds

Exposure threshold	Sensitivity	Specificity	PPV	NPV	Correct (%)
Exposure intensity in the current job					
0.2 (ACGIH)	80	81	87	71	80
0.5 (UK)	84	71	73	82	77
1	92	52	34	96	60
2	93	51	20	98	56
Average exposure intensity across all jobs					
0.2 (ACGIH)	76	77	84	65	76
0.5 (UK)	83	65	65	82	73
1	89	51	32	95	59
2	96	48	18	99	54

378 measure of exposure intensity or cumulative expo-  
 379 sure—whether in the current job, averaged across all  
 380 jobs, or as a measure of chronic exposure (leaving out  
 381 the last year). There were no meaningful differences  
 382 between our measures of chronic cumulative exposure  
 383 and intensity, and those measured across all jobs.  
 384 Lucchini et al. (1995) observed associations between  
 385 MnB and CEI across all jobs in workers who were not  
 386 currently exposed. In 1999, Lucchini et al. observed  
 387 associations between MnB and current exposure inten-  
 388 sity in currently exposed workers, but not with cumu-  
 389 lative exposure. Apostoli et al. (2000) and Lucchini  
 390 et al. (1995, 1999) propose that MnB reflects mainly  
 391 the body burden of Mn in currently unexposed workers,  
 392 and that this is why it is correlated with neurobehav-  
 393 ioural outcomes in such workers. For those currently  
 394 exposed, and for whom MnB reflects the impact of  
 395 current exposure as well as body burden, such correla-  
 396 tions are not found. Our results showing stronger  
 397 associations for current job measures, especially inten-  
 398 sity, provide some confirmation for Apostoli et al.'s  
 399 (2000) and Lucchini et al.'s (1995) observations that  
 400 the biomarkers of exposure significantly reflect current  
 401 atmospheric exposure intensity while workers are  
 402 exposed, as opposed to simply reflecting the body  
 403 burden of Mn when workers are removed from expo-  
 404 sure. Additionally, neither MnB or MnU in our study  
 405 were associated with neurobehavioural test results,  
 406 providing further confirmation of this interpretation  
 407 although this is not surprising as there were few con-  
 408 vincing neurobehavioural effects of Mn exposure  
 409 found in our study (Myers et al., in press). Had there  
 410 been clear effects it would have been possible to further  
 411 explore the utility of MnB might as a screening device.

412 At the individual level there were significant corre-  
 413 lations between log MnB and all other exposure vari-  
 414 ables—more strongly for intensity and cumulative  
 415 exposure in the current job. There were no meaningful

416 differences between correlations for atmospheric man-  
 417 ganese based on chronic measures or those based on  
 418 cumulative exposure across all jobs which is not  
 419 surprising since the mean years of service was high  
 420 at 17.2. MnB explained a relatively modest 33% of the  
 421 variance in atmospheric Mn. This was somewhat  
 422 higher than the 13% found by Apostoli et al. (2000)  
 423 who concluded that MnB is a poor biomarker of  
 424 exposure due to its high variability.

425 However, a ROC analysis was performed which  
 426 showed that it is possible to use an MnB cut-off of  
 427 10 µg/l to good effect as a screening tool to discrimi-  
 428 nate between individual exposures exceeding or falling  
 429 below a relatively strict atmospheric Mn exposure  
 430 threshold at the ACGIH TLV of 0.2 mg/m<sup>3</sup>. If the  
 431 Mn exposure threshold is calculated from exposure  
 432 intensity in the current job, an MnB exceeding 10  
 433 discriminates well for current exposure exceeding  
 434 the TLV. If the threshold is calculated from the average  
 435 intensity across all jobs MnB still discriminates well  
 436 between a body burden resulting from long-term expo-  
 437 sure intensity at the ACGIH TLV. For exposure inten-  
 438 sity in the current job, MnB exceeding 10 µg/l also  
 439 performs well at an exposure threshold of 0.5 mg/m<sup>3</sup>,  
 440 which is the exposure limit currently proposed by the  
 441 UK Health and Safety Executive (HSE). The value  
 442 10 µg/l usefully corresponds to the 95th percentile of  
 443 the MnB distribution in the unexposed referents.

444 While exceeding a MnB of 10 µg/l may be useful to  
 445 identify individuals in exceedance of strict thresholds  
 446 such as the ACGIH, its utility diminishes with more  
 447 lenient exposure thresholds due to increasingly poor  
 448 specificity. Similar analyses for MnU at a cut-off of  
 449 2.8 µg/l (95th percentile in the unexposed) at the  
 450 ACGIH TLV threshold showed poor discrimination  
 451 (sensitivity only 62% at specificity of 75%).

452 An important objective of the study was to identify  
 453 useful tools for surveillance and screening at group and

individual levels. Occupational hygiene surveys, and construction of complicated job exposure matrices, and linking exposure to health outcome data at a workplace are all demanding, time-consuming and expensive tasks. Surrogate measures for estimating exposure and risk are therefore very attractive. Of the three measures investigated here, MnB alone has utility for group exposure surveillance. The mean calculated from randomly sampled MnB measurements representative of a homogenous exposure group of workers with more than 1 year of service in their current job, could serve as a cross-sectional predictor of atmospheric Mn exposure. It could also serve well for surveillance of atmospheric Mn exposure intensity trends over time for that group/job, helping to monitor success in lowering workplace exposures. NIOSH sampling strategies could be employed (NIOSH, 1977).

At the individual level, it is possible to use an MnB of 10 µg/l cut-off to good effect as a screening tool discriminating well between individual exposures exceeding and falling below a relatively strict atmospheric Mn exposure range between the current ACGIH TLV of 0.2 mg/m<sup>3</sup> and the level currently being considered by the UK HSE (0.5 mg/m<sup>3</sup>). Positive predictive values will decline if exposure conditions improve in relation to this threshold range resulting in a declining prevalence of exceedances. MnU has poor utility as a measure of biological exposure as does serum prolactin as a measure of biological effect. These findings are not generalisable beyond those with similar exposures to manganese ferroalloy smelter workers.

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Cleveland (1979), Lucchini et al. (1997).

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